#### REVIEW



### Biomarkers of pyrethroid toxicity in fish

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#### Abstract

Pesticides contribute to human welfare by reducing vector-borne diseases and protecting crops against pests. Insecticides are the most widely employed pesticides for agricultural, domestic, and industrial pest control. However, some insecticides such as synthetic pyrethroids, analogs of the natural pyrethrin, persist in the environment and result in different hostile effects on nontarget organisms. Due to a continuous increase in the use of pyrethroids and their widespread application, different generations and types of pyrethroids have been frequently reported from environmental media, biota, and residential areas. Synthetic pyrethroids are observed to be less toxic to mammal and birds, relatively toxic to amphibians, and highly toxic to aquatic organisms including fish. Here, we review the occurrence, fate, biotransformation, and bioavailability of pyrethroids in waters. We also present biomarkers used to evidence toxicological effects of pyrethroids on fish. Toxic effects include oxidative stress and damage such as production of reactive oxygen species and lipid peroxidation; neurological behavioral inconsistencies; developmental effects such as delayed development and signaling; biochemical alterations of protein, glucose, and enzymes; hematological changes in white blood cells, red blood cells, and hemoglobin; physiological effects on metabolism and heart function; histopathological changes in the brain, liver, and gills; molecular toxicity including DNA damage, micronuclei induction, and altered gene or mRNA expression; and reproductive or endocrine disruption, e.g., disrupted pathways and signaling. Mechanisms of toxicity and control measures are also discussed.

Keywords Pesticides risk assessment  $\cdot$  Synthetic pyrethroids  $\cdot$  Oxidative stress  $\cdot$  Toxicological endpoints  $\cdot$  Multiple biomarkers  $\cdot$  Mechanism of action

#### Introduction

The human race made enormous progress; however, revolutionary achievements are coupled with environmental offputting factors such as vast use and release of drugs, heavy metals, fertilizers, and pesticides (Stankovic et al. 2014; Ullah et al. 2016a; Vieira et al. 2017; Afridi et al. 2018;

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Ali and Khan 2018; Khristoforova et al. 2018; Ullah and Li 2018). Pesticides are employed to repel, deter, or kill target organisms such as insects, algae, fungi, and bacteria in agricultural fields, public places such as homes, hospitals, and parks, industries, and public health programs (Ullah et al. 2019). With advancements in the field of pesticides chemistry, the numbers of pesticides are growing continuously. Different types of pesticides are used for targeting different types and/or species of organisms. The use of these pesticides is a major reason of elevating the standard of human life by different ways such as protecting the crops in the fields and stored food, destroying breeding site of different diseases causing insects, controlling harmful microorganisms including bacteria and viruses, and vanishing exasperating flies (Gill et al. 2018; Ullah et al. 2018a).

#### **Different classes of pesticides**

Pesticides are synthesized commercially and used under different names, belonging to different types and classes. The

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different classes of pesticides are employed based on their target organisms such as virucides against viruses, avicides against birds, algicides against algae, fungicides against fungi, nematicides against nematodes, rodenticides against rodents, herbicides against herbs, bactericides against bacteria, and acaricides or insecticides against insects (Regnery et al. 2018; Singh et al. 2018a; Valle et al. 2018). Among different classes of pesticides, insecticides are the most widely employed ones and attribute to about 80% of the use of the total pesticide (Ullah et al. 2018b). There are different registered classes of insecticides including organochlorines, carbamates, organophosphates, formamidines, organotins, organosulfurs, avermectins, neonicotinides, ryanodine, and rotenone, among others (Ullah et al. 2016b, c; Yang et al. 2018). However, one of the late introduced and most widely employed classes of insecticides is synthetic pyrethroids.

#### Introduction to pyrethroids

Pyrethroids are derived synthetically from pyrethrins, which are extracted from the flower of a plant, Chrysanthemum cinerariaefolium (Ullah 2015). Pyrethrins are insecticidal in nature due to the presence of ketoalcoholic esters of strongly lipophilic pyrethroic and chrysanthemic acids, having the capability of rapidly penetrating into insect bodies and leading to toxicosis. However, being highly sensitive to light natural pyrethrins break down within a few hours and cannot bioaccumulate in a sufficient concentration or amount to kill insects. With the help of modified structures, formulations, and stereochemistry, thousands of synthetic pyrethroids are developed. These modifications include cyano group addition, mixing of optical and geometric isomers, halogenation of the cyclopropane side chain of the pyrethrin molecule, adding different solvents and carriers, and different technical grade formulations (Kaviraj and Gupta 2014). These pyrethroids have a wide range of chemical and biological properties and performance; therefore, suitable pyrethroids are employed in agricultural fields, industries, parks, orchards, and homes (Ullah et al. 2018b).

### Biotransformation and environmental fate of synthetic pyrethroids

The routes for the elimination of synthetic pyrethroids in the environmental media include microbial degradation, photodegradation, volatilization, and hydrolysis (Gan et al. 2005). However, in the biological systems, pyrethroids are detoxified by two pathways—esterase-dependent hydrolytic reaction and oxidative reaction mediated by cytochrome P450s. The main factors recognized for nontarget organisms' susceptibility against synthetic pyrethroids are toxicokinetic factors. Synthetic pyrethroids are degraded through esterase-based hydrolysis followed by cytochrome P450sbased oxidation easily; therefore, they are relatively less toxic to mammals (Gammon et al. 2012). However, pyrethroids are highly toxic to fish because they lack hydrolase and therefore cannot swiftly detoxify synthetic pyrethroids hydrolytically like mammals (Yang et al. 2016). The only metabolic pathway of synthetic pyrethroids in fish is oxidative reaction catalyzed by cytochrome P450s. Different nonspecific metabolites of the synthetic pyrethroids have been recognized, such as 3-phenoxybenzoic acid, 3-phenoxybenzaldehyde, 3-phenoxybenzyl alcohol, 3-phenoxybenzoic acid, and 3-(2,2-dichlorovinyl)-2,2-dimethylcyclopropane carboxylic acid. The transformation and degradation of compounds in the environment depend upon their physicochemical properties (Singh et al. 2016). Synthetic pyrethroids have the property of hydrophobicity and are insoluble in water with an *n*-octanol distribution coefficient 6.6 in water. They are stable and persist in the aquatic sediments and soil (Gammon et al. 2012). Figure 1 shows some toxicological impacts, some metabolites, environmental fate or degradation, and biotransformation of synthetic pyrethroids.

### Bioavailability of synthetic pyrethroids in the aquatic environment

Synthetic pyrethroids lead to aquatic bodies through runoffs from the sprayed agricultural fields, parking lots, industries, and public health programs through spray drift to some extent; however, the main source of bioavailability in the water bodies is flowing therethrough rainstorm events. The magnitude and frequency of synthetic pyrethroids use and their precipitation patterns are observed to be critical factors governing synthetic pyrethroids transport to water bodies (Oros and Werner 2005). Moreover, the breakdown rates of the pyrethroids such as persistence on the soil surface, temperature, and canopy cover in association with their precipitation events may play a specific role in the determination of concentrations of synthetic pyrethroids in the runoffs (Palmquist et al. 2012). The concrete drainage system may transport a higher concentration of aqueous-phase pyrethroids in the urban and suburban areas as compared to earthen ditches channeled from agricultural particulate-rich runoffs (Weston and Lydy 2010).

#### Synthetic pyrethroids in the environment

Owing to the widespread applications of synthetic pyrethroids, these are reported from various parts of the world. Table 1 shows the reported concentrations of different synthetic pyrethroids from the soil, and land organisms or their products, whereas Table 2 shows the reported concentration of different synthetic pyrethroids from sediments, water,



Fig. 1 Synthetic pyrethroids (SPs) exposure leads to different toxicological impacts in the exposed organisms such as the induction of oxidative stress followed by oxidative damage, neurotoxicity, hematological toxicity, biochemical toxicity, and developmental toxicity. Some metabolites of synthetic pyrethroids have been identified, and their exposure led to different immunotoxicity, endocrine disruption, and reproductive toxicity. Pyrethroids get degraded by microbial deg-

fish, and other aquatic organisms from different countries across the globe. The range is given for individual synthetic pyrethroids, such that any of them observed in minimum and maximum concentrations. Synthetic pyrethroids are divided into two types, type I and type II. Type I pyrethroids are non-cyano pyrethroids, while type II pyrethroids contain the  $\alpha$ -cyano group. Figure 2 shows the chemical structure of widely employed esters of synthetic pyrethroids of different generations from both type I and type II groups, whereas Fig. 3 shows the chemical structure and formulae

radation, photodegradation, hydrolysis, and volatilization. Moreover, pyrethroids are biotransformed easily by mammals through hydrolytic (esterase) and oxidative (cytochrome P450s) reactions. Therefore, pyrethroids are less toxic to them. However, fish lack hydrolase and metabolize synthetic pyrethroids through oxidative (cytochrome P450s) reaction only. Therefore, they are highly toxic to fish and other aquatic organisms

of different esters of natural pyrethrin isolated from *C. cinerariaefolium*. Type II pyrethroids are considered to be more severely neuro-intoxicating as compared to type I, solely due to the presence of an  $\alpha$ -cyano group (Soderlund et al. 2002). Table 3 shows the acute toxic concentrations of type II synthetic pyrethroids against different fish species. Synthetic pyrethroids are widely used across the globe due to their low toxicity to mammals and birds. However, synthetic pyrethroids are known to pose marked hostile effects on aquatic organisms, more specifically on fish (Assis et al. 2009).

 Table 1
 Concentrations of pyrethroids from the soil, cash crops, and land/terrestrial organisms and/or their products across the globe

S. no.	Country	Detected pyrethroids	Conc. (min-max)	References
Soil (r	ng/g)			
1	China	Bifenthrin, fenpropathrin, cyhalothrin, cypermethrin, fenvalerate, and del- tamethrin	0.00-400.0	Liu et al. (2016)
2	Pakistan	Cypermethrin	0.07-1184.0	Rafique et al. (2016)
3	China	Bifenthrin and cyhalothrin	0.00-2.50	Zhu et al. (2015)
4	China	$\lambda$ -Cyhalothrin and cyfluthrin	7.50-38.00	Liu et al. (2015)
5	Togo	λ-Cyhalothrin	0.00-3.73	Mawussi et al. (2014)
6	India	Fenvalerate	0.00-19.17	Murugan et al. (2013)
7	China	Cypermethrin	5.00-36.00	Qiang et al. (2013)
8	China	Cypermethrin, cyhalothrin, and del- tamethrin	0.00-60.30	Yao et al. (2011)
9	USA	Bioallethrin and permethrin	1.97-724.19	Riederer et al. (2010)
10	India	Cypermethrin and fenvalerate	0.00-35.00	Kumari et al. (2008)
11	Spain	Permethrin, cypermethrin, deltame- thrin, $\lambda$ -cyhalothrin, and fenvalerate	4.00-60.00	Fernandez-Alvarez et al. (2008)
Cash	crops (ng/g)			
1	Singapore (vegetables)	β-Cyfluthrin, decamethrin, fenvalerate, permethrin, and bifenthrin	4.17 (Avg.)	Yu and Yang (2017)
2	China (fruits: apple, pear, peach, and grape)	λ-Cyhalothrin, cypermethrin, fen- propathrin, fenvalerate, bifenthrin, cyfluthrin, and deltamethrin	5.00-1208.00	Li et al. (2016)
3	China (jujubes and persimmon)	Bifenthrin, fenpropathrin, cyhalothrin, fenvalerate, deltamethrin, and cyper- methrin	59.40–2945.0	Liu et al. (2016)
4	China (dates)	Bifenthrin, fenpropathrin, fenvalerate, and $\lambda$ -cyhalothrin	3.30-1100.00	Yun et al. (2015)
5	Brazil (sweet pepper)	Amethrin and fenpropathrin	53.60 (Avg.)	Kemmerich et al. (2015)
6	Poland (apple, orange, and cabbage)	Cypermethrin and bifenthrin	20.00-300.00	Szpyrka (2014)
7	Ghana (eggplant, okra, and tomato)	Allethrin, permethrin, and deltamethrin	3.00-133.00	Akoto et al. (2015)
8	China (vegetables)	Cyhalothrin, bifenthrin, cypermethrin, permethrin, fenvalerate, decamethrin, fenpropathrin, and cyfluthrin	8.00-260.00	Li et al. (2014)
9	Ghana (okra)	Bifenthrin, $\lambda$ -cyhalothrin, permethrin, cyfluthrin, cypermethrin, fenvalerate, and deltamethrin	100.00-4100.00	Essumang et al. (2013)
10	Thailand (fruits/vegetables)	Tetramethrin, fenpropathrin, cyperme- thrin, deltamethrin, fenvalerate, and permethrin	10.00–770.00	Wongsa and Burakham (2012)
11	Mexico (vegetables)	Cyhalothrin, cyfluthrin, cypermethrin, fenvalerate, and deltamethrin	4.00–573.00	Aldana-Madrid et al. (2011)
12	Philippines (eggplant)	Cypermethrin	10.00-30.00	Lu (2011)
Land/	terrestrial organisms and/or their products	(ng/g and/or ng/L)		
1	France (Partridge (Perdix perdix) eggs)	λ-Cyhalothrin	$0.34 \times 10^{-6}$	Bro et al. (2016)
2	India (bovine milk)	Cyhalothrin, cypermethrin, fenvalerate, and deltamethrin	0.50-0.90	Bedi et al. (2015)
3	Egypt (honeybee)	Cyhalothrin, fenvalerate, and perme- thrin	1.30-2.70	Malhat et al. (2015)
4	Egypt (honey)	Cyhalothrin, permethrin, and fenvaler- ate	1.30–19.00	Malhat et al. (2015)
5	Brazil (honey from: Apis mellifera)	Bifenthrin	0.28-2.50	Pacífico da Silva et al. (2015)
6	Brazil (honey from: <i>Melipona sub-</i> <i>nitida</i> )	Bifenthrin	0.092-2.50	Pacífico da Silva et al. (2015)

Table 1 (continued)

S. no.	Country	Detected pyrethroids	Conc. (min-max)	References
7	France (honey)	Cypermethrin, deltamethrin, esfenvaler- ate, bifenthrin, permethrin, resme- thrin, and λ-cyhalothrin	0.20-70.00	Paradis et al. (2014)
8	India (frog species: Fejervarya limno- charis and Haplobatrachus crassus)	Esfenvalerate	26.91-42.15	Kittusamy et al. (2014)
9	China (mutton)	Tetramethrin and cyhalothrin	2.10-114.50	Gao et al. (2012)
10	China (acacia honey)	Etofenprox, meperfluthrin, cyperme- thrin, and tetramethrin	0.00-2.60	Zhang et al. (2011)
11	China (vitex honey)	Etofenprox, meperfluthrin, cyperme- thrin, and tetramethrin	0.00-3.90	Zhang et al. (2011)
12	Brazil (female dogs: Canis lupus)	Allethrin, cyhalothrin, cypermethrin, deltamethrin, and tetramethrin	2000.00-55,000.00	Andrade et al. (2010)
13	USA (honey and honeybee)	Allethrin, cypermethrin, bifenthrin, cyfluthrin, cyhalothrin, deltamethrin, fluvalinate, flumethrin, permethrin, prallethrin, and tefluthrin	24.00–19,600.00	Johnson et al. (2010)

Keeping in view the toxic effects of synthetic pyrethroids on fish, different biomarkers are used to delineate their toxic impacts as well as envisaging biomarkers for future research.

#### Biomarkers of pyrethroid toxicity in fish

Biomarkers are indicators of the response of exposure to any toxicant, chemical, pollutant, or any other foreign particle. Biomarkers can be evaluated at a molecular or cellular level to community or ecosystem level. These biomarkers can substantially reveal the toxic effects of the toxicant on the exposed organism, such as toxicities on their neurology resulting in altered behavior, histopathological, morphological, anatomical, physiological, hematological, and biochemical profiles. Table 4 presents various toxic effects of pyrethroids to biomarkers in different fish species.

### Pyrethroid-induced oxidative stress or damage in fish

Oxidative stress is widely employed as a sensitive biomarker in ecotoxicological assessments in order to understand the underlying hostile effects. The oxidative stress is evaluated in terms of reactive oxygen species (ROS) or free radicals' production, increased lipid peroxidation, and altered activities of the antioxidant enzymes in response. ROS production leads to oxidative damage at the cellular level to DNA, lipids, and protein (Ullah et al. 2018a). To cope with the oxidative damage and to defend the cell against free radicals, different stress proteins such as heat-shock proteins, glucoseregulating proteins, and antioxidant enzymes including catalase, peroxidases, superoxide dismutase, glutathione reductase, glutathione-S-transferase, and glutathione peroxidase are produced. However, when the production of the free radicals exceeds the potential of the defense system of the exposed organisms, it leads to different levels of oxidative damage such as DNA damage (Ullah et al. 2017). Research revealed that synthetic pyrethroids-induced oxidative stress leads to the different type of instant toxicities as well as toxicities and weak immunity at later stages in fish. In response to the oxidative stress, the fish adapt defensive mechanism by changing their antioxidant enzymatic activities such as increasing their activities to cope with the free radicals.

Exposure to different synthetic pyrethroids induced oxidative stress in different species of fish, for example, cypermethrin induced oxidative stress in different tissues of *Tor putitora* (Ullah et al. 2014), *Labeo rohita* (Ullah 2015), and *Oncorhynchus mykiss* (Kutluyer et al. 2015), cyhalothrin induced oxidative stress in different tissues of *Cyprinus carpio* (Clasen et al. 2018) and *Prochilodus lineatus* (Vieira and dos Reis Martinez 2018), and deltamethrin induced oxidative stress in different tissues of *Cyprinus carpio* (Ensibi et al. 2013), *Sparus aurata* (Guardiola et al. 2014), *Oreochromis niloticus* (Abdel-Daim et al. 2015), *Danio rerio* (Parlak 2018), and *Hypophthalmichthys molitrix* (Ullah et al. 2019).

#### Neurotoxicity

Pyrethroids exert toxic effects on the nervous system of the fish by affecting their sodium channels. They attached to these gated channels and delay the inactivation of the Na<sup>+</sup> channels, which ultimately led to neuronal excitability (Ullah et al. 2019). However, recent research revealed that synthetic pyrethroids also affect the other voltagegated channels such as calcium and chloride channels, and receptor of  $\gamma$ -aminobutyric acid as their secondary targets

ou v	Country	Deterted nursthroids	Conc (min_max)	References
		proceed by removed		
Sedin	nents (ng/g)			
	Mexico	Bifenthrin, cypermethrin, permethrin, cyfluthrin, cyhalo- thrin, esfenvalerate, fenpropathrin, and deltamethrin	0.08–1.97	Hernández-Guzmán et al. (2017)
6	China	Bifenthrin, cyfluthrin, cypermethrin, λ-cyhalothrin, deltamethrin, esfenvalerate, fenpropathrin, Permethrin, and tefluthrin	0.00–29.80	Sun et al. (2016)
3	USA	Tralomethrin, permethrin, fenpropathrin, λ-cyhalothrin, cypermethrin, deltamethrin, esfenvalerate, bioresme- thrin, cyfluthrin, and bifenthrin	27.15-1407.63	Chinen et al. (2016)
4	China	Bifenthrin, cyfluthrin, cypermethrin, deltamethrin, esfen- valerate, fenpropathrin, permethrin, and tefluthrin	1.70–1631.00	Sun et al. (2016)
5	Nigeria	λ-Cyhalothrin and permethrin	770.00-375,700	Olutona et al. (2016)
9	China	Esfenvalerate and permethrin	0.00–991.00 and 0.00–562.00	Zhao et al. (2016)
Ζ	Australia	Bifenthrin, permethrin, and cypermethrin	< 2.00-930.00	Marshall et al. (2016)
8	China	Fenpropathrin, bifenthrin, permethrin, cyfluthrin, etofenprox, fenvalerate, deltamethrin, cyhalothrin, and cypermethrin	1.00-20.00	Zheng et al. (2016)
6	Pakistan	Deltamethrin	660.00-5580.00	Mahboob et al. (2015)
10	China	Cypermethrin, esfenvalerate $\lambda$ -cyhalothrin, and permethrin	0.11-0.21	Qi et al. (2015)
11	Pakistan	Cypermethrin and deltamethrin	183.00-318.00	Jabeen et al. (2015)
12	China	Cypermethrin	50.00-4960.00	Sun et al. (2015)
13	Vietnam	Permethrin	0.00-59,741.00	Duong et al. (2014)
14	Thailand	Cypermethrin	0.20-52.90	Sangchan et al. (2014)
15	Turkey	Etofènprox	9.50-385.00	Yurtkuran and Saygı (2013)
16	China	Fenpropathrin, A-cyhalothrin, permethrin, cypermethrin, fenvalerate, biphenyl pyrethroid, and deltamethrin	0.00-94.00	Zhao et al. (2013)
17	USA	Bifenthrin, permethrin, deltamethrin, and cypermethrin	2.10 - 123.00	Weston et al. (2011)
18	China	Permethrin, λ-cyhalothrin, bifenthrin, fenpropathrin, fenvalerate, cypermethrin, cyfluthrin, and deltamethrin	0.00-2.91	Fang et al. (2012)
19	USA	Bifenthrin, cyfluthrin, permethrin, cypermethrin, deltamethrin, esfenvalerate, fenpropathrin, and λ-cyhalothrin	0.00–132	Lao et al. (2012)
20	USA	Bifenthrin, cyhalothrin, cypermethrin, permethrin, and resmethrin	0.10-38.00	Kuivila et al. (2012)
21	China	Bifenthrin, cyfluthrin, fenpropathrin, λ-cyhalothrin, cypermethrin, deltamethrin, esfenvalerate, and perme- thrin	0.65–384.00	Li et al. (2011)

Table	2 (continued)			
S. no.	Country	Detected pyrethroids	Conc. (min-max)	References
22	USA	Allethrin, bifenthrin, cyfluthrin, $\lambda$ -cyhalothrin, cyper- methrin, deltamethrin, esfenvalerate, fenpropathrin, fluvalinate, permethrin, sumithrin, and tetramethrin	0.00-217.00	Domagalski et al. (2010)
23	China	Tetramethrin, flucythrinate, $\beta$ -cyfluthrin, fenpropathrin, $\lambda$ -cyhalothrin, $\alpha$ -cypermethrin, deltamethrin, fenvaler- ate, bioresmethrin, tau-fluvalinate, permethrin, and bifenthrin	5.60–9.50	Yang et al. (2010)
24	Spain	Deltamethrin and cypermethrin	8.27-71.90	Feo et al. (2010b)
Water	r(ng/L)			
1	Greece	Alphamethrin, cypermethrin, esfenvalerate, λ-cyhalothrin, and permethrin	0.00-259.00	Papadakis et al. (2018)
7	China	Fenpropathrin, bifenthrin, permethrin, cyfluthrin, etofenprox, fenvalerate, deltamethrin, cyhalothrin, and cypermethrin	<100.00-1388.00	Zheng et al. (2016)
б	Sudan	λ-Cyhalothrin, cypermethrin, fenvalerate, and deltame- thrin	0.00-926.50	Nesser et al. (2016)
4	USA	Bifenthrin and esfenvalerate	0.00-120.00	Carpenter et al. (2016)
5	China	Cypermethrin	12,470–19,080	Chen et al. (2015)
9	Bangladesh	Cypermethrin	0.00-80,500.00	Hossain et al. (2015)
٢	Pakistan	Cypermethrin	4.20-2680.00	Mahboob et al. (2015)
8	Vietnam	Cypermethrin	770.00	Chau et al. (2015)
6	Greece	Alphamethrin, permethrin, and $\lambda$ -cyhalothrin	0.00-161.00	Papadakis et al. (2015)
10	USA	Bifenthrin, cyfluthrin, cypermethrin, cyhalothrin, and permethrin	0.00–27.00	Weston et al. (2014)
11	Czech Republic	Flumethrin, fenpropathrin, tau-fluvalinate, and fenvaler- ate	26,110 (River), 34,530 (lake)	Fernández-Ramos et al. (2014)
12	China	Permethrin, tefluthrin, bioallethrin, prallethrin, benzylp- ermethrin, cyhalothrin, meperfluthrin, fenvalerate, deltamethrin, and pyrethrin	0.00-13.90	Liu and Zhang (2014)
13	China	Permethrin, bifenthrin, and cypermethrin	4680.00-29,720.00	Wang et al. (2013)
14	USA	Bifenthrin, cypermethrin, permethrin, and cyfluthrin	0.00-20.00	Ensminger et al. (2013)
15	China	Cyphenothrin	52.00	Li and Chen (2013)
16	Turkey	Etofenprox	1.00-8.00	Yurtkuran and Saygı (2013)
17	Mexico	Cyhalothrin and cypermethrin	29,400.00	Moreno-Villa et al. (2012)
18	Philippines	Cyhalothrin, deltamethrin, and cypermethrin	0.50-1400.00	Elfman et al. (2011)
19	China	Biphenyl pyrethroid, fenpropathrin, cyhalothrin, fenvaler- ate, and deltamethrin	0.00-6360.00	Wang et al. (2010)
20	Spain	Cypermethrin and deltamethrin	0.73-58.78	Feo et al. (2010b)
21	Spain	Cypermethrin	4.94–30.50	Feo et al. (2010a)

Table	2 (continued)			
S. no.	Country	Detected pyrethroids	Conc. (min-max)	References
22	India	Cypermethrin	17.00–90.00	Kumari et al. (2008)
23	USA	Allethrin, bifenthrin, cyfluthrin, cypermethrin, delta/ tralomethrin, fenpropathrin, fenvalerate, flucythrinate, $\lambda$ -cyhalothrin, permethrin, pyrethrin, phenothrin, pral- lethrin, resmethrin, and tetramethrin	1.71–12.8	Woudneh (2006)
24	USA	Permethrin, esfenvalerate, $\lambda$ -cyhalothrin, and bifenthrin	200.00-500.00	Weston et al. (2004)
Fish (i	ug/g)			
1	Spain (Salmo salar)	Bifenthrin, cyfluthrin, cyhalothrin, cypermethrin, del- tamethrin, fenvalerate, permethrin, and tetramethrin	0.10–25.00	Aznar-Alemany et al. (2019)
5	Norway (Salmo salar)	Tetramethrin, bifenthrin, cyhalothrin, permethrin, cyfluthrin, cypermethrin, deltamethrin, and fenvalerate	0.00-4.42	Aznar-Alemany et al. (2017)
ŝ	Scotland (Oncorhynchus mykiss)	Bifenthrin, cyhalothrin, permethrin, cypermethrin, and deltamethrin	0.00-1.83	Aznar-Alemany et al. (2017)
4	Denmark (Oncorhynchus mykiss)	Bifenthrin and permethrin	0.05-0.44	Aznar-Alemany et al. (2017)
5	France (Salmo salar)	Bifenthrin, permethrin, cypermethrin, and deltamethrin	0.00-0.24	Aznar-Alemany et al. (2017)
9	Spain (Oncorhynchus mykiss and Salmo salar)	Cypermethrin and deltamethrin	0.00-0.56	Aznar-Alemany et al. (2017)
٢	Chile (Oncorhynchus mykiss, O. kistuch, O. goroscha, and O. keta)	Bifenthrin, cyhalothrin, cypermethrin, and deltamethrin	0.00-0.28	Aznar-Alemany et al. (2017)
×	Brazil (frozen fish muscle)	Imiprothrin, resmethrin, tetramethrin, phenothrin, cyhalothrin, permethrin, fenvalerate, cyfluthrin, and cypermethrin	0.00-26.10	Dallegrave et al. (2016)
6	China (Mugil cephalus)	Fenpropathrin, bifenthrin, permethrin, cyfluthrin, etofenprox, fenvalerate, deltamethrin, cyhalothrin, and cypermethrin	<100.00-1388.00	Zheng et al. (2016)
10	Spain (L. graellsii, L. Sclateri, B. guiraonis, Cyprinus carpio, Salmo trutta, Silurus glanis, and Gn lozanoi)	Bifenthrin, permethrin, cyfluthrin, cyhalothrin, and cypermethrin	12.00-4938.00	Corcellas et al. (2015)
11	Pakistan ( <i>Labeo rohita</i> )	Cypermethrin	52.00-283.00	Jabeen et al. (2015)
12	China ( <i>Ctenopharyngodon idella</i> )	Cypermethrin	3.27-74.29	Chen et al. (2015)
13	Pakistan (Cyprinus carpio)	Deltamethrin	490.00-839.00	Jabeen et al. (2015)
14	Pakistan ( <i>Catla catla</i> )	Deltamethrin and cypermethrin	350.00-3470.00	Mahboob et al. (2015)
15	China (Tilapia mossambica)	Cypermethrin	8.57-13.41	Wang (2013)
16	China (Carassius auratus and Mugil cephalus)	Cyfluthrin, cypermethrin, and deltamethrin	0.00-4.68	Wang et al. (2012)
17	China (Carassius carassius)	Phenothrin and $\alpha$ -tetramethrin	0.30-416.23	Cai et al. (2011)
18	India (brain: Bagarius bagarius)	Cypermethrin and deltamethrin	0.19-4.87	Vengayil et al. (2011)
19	India (liver: Bagarius bagarius)	Cypermethrin and deltamethrin	0.35-6.68	Vengayil et al. (2011)
20	India (ovary: Bagarius bagarius)	Cypermethrin and deltamethrin	0.15–2.30	Vengayil et al. (2011)
21	Canada (Salmo salar, Oncorhynchus mykiss, Tilapia sp.)	Cypermethrin and deltamethrin	0.30-6.50	Rawn et al. (2010)

S. no.	Country	Detected pyrethroids	Conc. (min-max)	References
Other ,	aquatic organisms (ng/g)			
1	Mexico (mussel: Mytilus californianus)	Permethrin, cypermethrin, and bifenthrin	1.70-4.90	Hernández-Guzmán et al. (2017)
5	USA (Hyalella azteca)	Bifenthrin, cyhalothrin, cypermethrin, permethrin, and resmethrin	0.50-11.2	Kuivila et al. (2012)
$\mathfrak{S}$	Brazil (dolphins' liver: Pontoporia blainvillei)	Fenvalerate, tetramethrin, bifenthrin, tetramethrin, fluvalinate, <i>λ</i> -cyhalothrin, permethrin, cyfluthrin, and cypermethrin	0.18–18.9 (Adults) 0.10–11.0 (Juveniles) 0.10–31.6 (Calves)	Alonso et al. (2012)
4	Brazil (dolphins' breast milk: Pontoporia blainvillei)	Fenvalerate, tetramethrin, bifenthrin, tetramethrin, fluvalinate, <i>λ</i> -cyhalothrin, permethrin, cyfluthrin, and cypermethrin	2.5-4.8 (Total SPs conc.)	Alonso et al. (2012)
5	Brazil (dolphins' placenta: Pontoporia blainvillei)	Fenvalerate, tetramethrin, bifenthrin, tetramethrin, fluvalinate, λ-cyhalothrin, permethrin, cyfluthrin, and cypermethrin	331.0-1812.0 (Total SPs conc.)	Alonso et al. (2012)

Table 2 (continued)

(Soderlund 2012). Disturbance to these channels leads to different neurobehavioral changes. Moreover, the neurotoxic effects lead to complex consequences such as affected energy metabolism, neuromuscular functions, neural transduction, and homeostasis. Figure 2 shows the neurotoxic effects of synthetic pyrethroids, their mechanisms, and subtle consequences. The neurotoxic effects can be in the form of disturbed voltage-gated channels, behavioral inconsistencies or alterations, and inhibition of acetylcholinesterase activity.

#### Behavioral inconsistencies and alterations

Research revealed that exposure to synthetic pyrethroids resulted in different behavioral inconsistencies in fish, such as sluggish movement, disturbed swimming or swimming pattern, inability to maintain their position, reduced feeding, interrupted school behavior, hypo- or hyperexcitability, dangling or irregular or erratic swimming, increased opercula movements, rapid jerky movements, loss of equilibrium, frequently surfacing, adapting vertical position, sinking to bottom, hypo- or hyperactiveness, jumping, loss of balance, motionlessness, and disturbed migratory pattern in different fish species such as Tor putitora (Ullah et al. 2014), Labeo rohita (Ullah 2015), and Clarias batrachus (Kumar et al. 2011b). The acetylcholinesterase is active at both the neural and neuromotor junctions of the muscle tissues; therefore, the neuromuscular inhibition of acetylcholinesterase leads to blocked neural transmission and increased acetylcholine at the nerve endings, which consequently lead to different behavioral inconsistencies. Therefore, these alterations are often associated with the inhibition of acetylcholinesterase activity in the brain or muscles of the fish and/or increment in the level of acetylcholine.

#### Inhibition of acetylcholinesterase activity

A number of research studies revealed that synthetic pyrethroids induce neurotoxic effects by inhibiting the activity of acetylcholinesterase or incrementing the level of acetylcholine in the brain of various fish tissues. The inhibition of acetylcholinesterase results in nerve impulses and makes them permeable to sodium. Synthetic pyrethroids delay the closing of sodium channels, allowing sodium inflow in a heavy concentration, which consequently leads to multiple never impulses, which in turn release a neurotransmitter, acetylcholine, leading to their higher accumulation in the nerve synapses and ultimately decreased cholinergic transmission and other neurotoxic effects. In fish, these effects are increased operculum movement, convulsions, and surfacing (Singh et al. 2018b). Deltamethrin exposure resulted in inhibition of the acetylcholinesterase in the brain of silver carp resulting in erratic swimming, vertical position adaptation, hyperactivity, and equilibrium loss as well as in the muscle



Fig.2 Synthetic pyrethroids (SPs) induced neurotoxicity (mechanism and subtle consequences) and chemical structures of different SPs [type I (without  $\alpha$ -cyano group) including allethrin (first generation), permethrin (second generation), resmethrin (third generation),

tissues that resulted in the desensitization of the receptors of nicotine acetylcholine and subsequently resulted in muscular weakness and changed swimming pattern (Ullah et al. 2019).

#### **Developmental toxicity**

Synthetic pyrethroids are reported to exhibit greater acute toxic effects on the developing stages of animals as compared to adult stages (Yang et al. 2018). However, fish is highly sensitive and more susceptible to synthetic pyrethroids during their early life stages as compared to their adult stage (Yang et al. 2014). Synthetic pyrethroids also have the capability of affecting the development and growth of various animals (DeMicco et al. 2010). There is a continuously growing body of evidence, revealing the developmental toxicity of different pyrethroids on and bifenthrin (fourth generation), while type II (with  $\alpha$ -cyano group) including fenvalerate (third generation), cyhalothrin (fourth generation), cypermethrin (fourth generation), and deltamethrin (fourth generation)]

nontarget organisms, more specifically against fish, for example, exposure to bifenthrin accelerated hatching and impaired the normal morphology of Danio rerio, same as by cypermethrin, by inducing craniofacial abnormalities, pericardial edema, body curvatures, yolk edema, and crooked body (DeMicco et al. 2010; Jin et al. 2009; Shi et al. 2011). Similarly, joint exposure of Danio rerio to cypermethrin and permethrin led to different toxicities at larval stage (Yang et al. 2014), bifenthrin disturbed the dopaminergic signaling at the juvenile stage of Oncorhynchus mykiss (Crago and Schlenk 2015), cypermethrin induced different developmental deformities and altered the enzymatic activities in the developmental stages of Labeo rohita (Dawar et al. 2016), and deltamethrin induced oxidative stress leading to apoptosis and different morphological alterations in *Danio rerio* (Parlak 2018).



Fig. 3 The isolated pyrethrins from *Chrysanthemum cinerariaefolium* flowers with their chemical structures and formulae—the shown pyrethrins are esters of natural pyrethrum. Mechanism of action of synthetic pyrethroids (SPs): (1) SPs exposure results in the production of excessive reactive oxygen species (ROS) that leads to detrimental effects on lipids, proteins, and DNA, (2) SPs interrupt ligand-gated channel and allow inflow of sodium in a higher concentration which leads to multiple nerve impulses and ultimately to inhibition of acetylcholinesterase and accumulation of acetylcholine, which stimulate other nerves, (3) SPs affect voltage-gated calcium chan-

#### Hematological toxicity

Hematology is often assessed as a useful biomarker in eco-, aquatic, pesticides, and fisheries toxicology. Synthetic pyrethroids exposure results in different hematotoxic effects because after entering into the fish body, blood and blood-producing hematopoietic tissues are continuously exposed to the destructive effects of the respective pyrethroid. Exposure of *Tor putitora* to the acute concentration of cypermethrin led to an increase in white blood cells and a decrease in red blood cells (Ullah et al. 2015). Similarly, a number of studies reported different types of toxic effects on the hematological profile including white blood cells such as lymphocytes, thrombocytes, granulocytes, and monocytes, red blood cells, hemoglobin, packed cell volume, mean corpuscular volume, mean corpuscular hemoglobin concentration, and

nel and consequently (4) increase calcium concentration in the cytosol that consequently lead to cytotoxicity, (5) SPs inhibit receptor of  $\gamma$ -aminobutyric acid that consequently inhibits GABA receptor and ultimately leads to excitability and convulsion, (6) SPs disturb ATP formation/synthesis directly as well as glucose regulation is disturbed in response to cortisol regulation, (7) SPs exposure leads to the retention of acetylcholine in the synaptic gap (due to closing of synaptic cleft) which increases acetylcholine level, and (8) SPs lead to genotoxicity either by oxidative stress induction or due to accumulation of calcium in a higher concentration in the cytosol

mean corpuscular hemoglobin of different fish species after exposure to different synthetic pyrethroids such as Cyprinus carpio (Velisek et al. 2009a), Catla catla (Vani et al. 2012), Rhamdia quelen (Montanha et al. 2014), and Alburnus tarichi (Özok et al. 2018) in response to cypermethrin, Catla catla (Vani et al. 2011) and Salmo trutta fario (Karatas 2016) in response to deltamethrin, and Prochilodus lineatus in response to  $\lambda$ -cyhalothrin (Vieira and dos Reis Martinez 2018). The alterations in the hematological parameters including red blood cells might be attributed to the inhibition of hemosynthesis or erythropoiesis, destruction of blood cells such as red blood cells (anemia), decreased genesis of the red blood cells due to hypoxia, less hemoglobin or no hemoglobin, hematopoietic system's failure, and osmoregulatory dysfunction, whereas white blood cells may be altered due to the stimulated defense mechanism or immune system

Table 3Acute toxic concentrations ( $LC_{50}$ ) of different type II synthetic pyrethroids against different fish species

S. no.	Fish species	LC <sub>50</sub>	References
Deltame	ethrin		
1	Hypophthalmichthys molitrix	2.0 μg/L (96 h)	Ullah et al. (2019)
2	Colossoma macropomum	55.0 μg/L (96 h)	Cunha et al. (2018)
3	Channa punctatus	7.33 μg/L (96 h)	Singh et al. (2018b)
4	Ctenopharyngodon idella	0.331 μg/L (96 h)	Srinivasa Rao et al. (2018)
5	Clarias gariepinus	51.89 μg/L (96 h)	Hamed (2016)
6	Salmo trutta fario	4.7 μg/L (96 h)	Karatas (2016)
7	Labeo rohita	0.38 mg/L (96 h)	Suvetha et al. (2015)
8	Cirrhinus mrigala	8.0 mg/L (96 h)	David et al. (2015)
9	Anabas testudineus	0.07 mg/L (96 h)	Sapana Devi and Gupta (2014)
10	Channa punctatus	0.75 mg/L (96 h)	Jayaprakash and Shettu (2013)
11	Clarias gariepinus	0.75 µg/L (48 h)	Amin and Hashem (2012)
12	Xiphophorus helleri	2.87 μg/L (96 h)	Khalili et al. (2012)
13	Oreochromis niloticus	1.70 µg/L (48 h)	Kan et al. (2012)
14	Catla catla	4.84 μg/L (96 h)	Vani et al. (2011)
15	Oncorhynchus mykiss	0.3 and 0.6 µg/L	Atamanalp and Erdoğan (2010)
16	Heteropneustes fossilis	1.5 mg/L (96 h)	Srivastav et al. (2010)
17	Puntius chrysopterus	14.2 µg/L (96 h)	Pawar et al. (2009)
18	Labeo rohita	1.00 mg/L (96 h)	Rathnamma et al. (2009)
19	Danio rerio	0.5–1.0 μg/L (96 h)	Koc et al. (2009)
20	Poecilia reticulata	19.0 μg/L (96 h)	Stalin et al. (2008)
21	Oreochromis niloticus	14.6 μg/L (96 h)	El-Sayed and Saad (2008)
22	Oncorhynchus mykiss	20 μg/L (96 h)	Velíšek et al. (2007)
23	Carassius auratus	2.0 μg/L (48 h)	Costin et al. (2007)
24	Silurus glanis	0.686 μg/L (48 h)	Köprücü et al. (2006)
25	Oreochromis niloticus	4.85 μg/L (48 h)	Yildirim et al. (2006)
26	Oncorhynchus mykiss	0.7 μg/L (96 h)	Ural and Sağlam (2005)
27	Cyprinus carpio	0.074 µg/L (48 h)	Köprücü and Aydın (2004)
Cyperm	ethrin		
1	Cirrhinus mrigala	0.85 µg/L (96 h)	Singh (2017)
2	Oreochromis niloticus	4.85 (Commercial) and 9.74 (technical grade) $\mu g/L$ (96 h)	Majumder and Kaviraj (2017)
3	Heteropneustes fossilis	0.075 ml/L (96 h)	Monir et al. (2016)
4	Cnesterodon decemmaculatus	1.89–2.60 µg/L (96 h)	Brodeur et al. (2016)
5	Danio rerio	1.94 (10-dpf) and 3.56 (20-dpf) µg/L (96 h)	Rodríguez-Estrada et al. (2016)
6	Oreochromis niloticus	6.2 μg/L (96 h)	Haque and Mondal (2016)
7	Labeo rohita	5.0 μg/L (96 h)	Ullah (2015)
8	Tor putitora	63.0 μg/L (96 h)	Ullah et al. (2015)
9	Heteropneustes fossilis	0.67–1.27 μg/L (72 h)	Saha and Kaviraj (2013)
10	Heteropneustes fossilis	3.783 μg/L (96 h)	Bhutia et al. (2013)
11	Catla catla	4.43 μg/L (96 h)	Vani et al. (2012)
12	Clarias gariepinus	0.25 μg/L (96 h)	Akinrotimi et al. (2012)
13	Cirrhinus mrigala	150.0 μg/L (96 h)	Vasantharaja et al. (2012)
14	Hypophthalmichthys molitrix	0.917 μg/L (96 h)	Shaluei et al. (2012)
15	Rutilus rutilus caspicus	0.627 μg/L (96 h)	Shaluei et al. (2012)
16	Cyprinus carpio	250.0 µg/L (96 h)	Meenambal et al. (2012)
17	Labeo rohita	0.205 µg/L (24 h)	Tiwari et al. (2012)
18	Clarias batrachus	12.0 µg/L (96 h)	Kumar et al. (2011a)
19	Danio rerio	0.29 (48 h) and 0.27 mg/L (72 h)	Xu et al. (2010)
20	Colisa fasciatus	6.0 μg/L (96 h)	Singh et al. (2010)
21	Clarias batrachus	0.21 mg/L (96 h)	Begum (2009)

#### Table 3 (continued)

S. no.	Fish species	LC <sub>50</sub>	References
22	Oreochromis niloticus	5.99 µg/L (96 h)	Sarıkaya (2009)
23	Oryzias latipes	30.8 (Adult), 38.5 (fry), and 111.4 (embryo) µg/L (48 h)	Kim et al. (2008)
24	Clarias gariepinus	0.063 mg/L (96 h)	Ayoola and Ajani (2008)
25	Rhamdia quelen	0.193 mg/L (96 h)	Borges et al. (2007)
26	Channa punctatus	0.4 mg/L (96 h)	Kumar et al. (2007)
27	Cyprinus carpio	0.256–5.074 μg/L (48 h)	Aydın et al. (2005)
28	Heteropneustes fossilis	7.20 µg/L (96 h)	Mishra et al. (2005)
29	Poecilia reticulata	9.43 µg/L (96 h)	Yılmaz et al. (2004)
Fenvale	rate		
1	Anabas testudineus	472.5 (Static) and 376.0 (Cont. flow) mg/L (96 h)	Satyavardhan (2013)
2	Puntius sophore	1.789 (Static) and 1.415 (Cont. flow) mg/L (96 h)	Satyavardhan (2013)
3	Cyprinus carpio	2.171 (Static) and 1.775 (Cont. flow) mg/L (96 h)	Satyavardhan (2013)
4	Ctenopharyngodon idellus	2.627 (Static) and 2.121 (Cont. flow) mg/L (96 h)	Satyavardhan (2013)
5	Clarias batrachus	1.35 µg/L (96 days)	Datta and Kaviraj (2011)
6	Heteropneustes fossilis	0.65 µg/L (96 days)	Datta and Kaviraj (2011)
7	Labeo rohita	5.36 µg/L (96 h)	Prusty et al. (2011)
8	Danio rerio	6.25 µg/L (96 h)	Gu et al. (2010)
9	Cyprinus carpio	3.059 µg/L (96 h)	Raja et al. (2010)
10	Danio rerio	8.29 µg/L (96 h)	Ma et al. (2009)
11	Clarias gariepinus	4.24–2.94 µg/L (96 h)	Bhattacharya and Kaviraj (2009)
12	Channa punctatus	2.13 µg/L (96 h)	Singh et al. (2007)
13	Catla catla	6.0 µg/L (96 h)	Tandon et al. (2005)
14	Clarias gariepinus	250 µg/L (48 h)	Sakr et al. (2005)
15	Cirrhinus mrigala	6.0 µg/L (96 h)	Mushigeri and David (2004)
Cyhalot	hrin		
1	Prochilodus lineatus	5.0 μg/L (<24 h)	Vieira and dos Reis Martinez (2018)
2	Brycon amazonicus	0.65 µg/L (96 h)	Venturini et al. (2018)
3	Gambusia affinis	1.107 µg/L (96 h)	Güner (2016)
4	Cyprinus carpio	0.160 µg/L (96 h)	Bibi et al. (2014)
5	Labeo rohita	0.7 µg/L (96 h)	Dey and Saha (2014)
6	Oreochromis niloticus	2.901 µg/L (96 h)	Piner and Üner (2012)
7	Clarias batrachus	5.0 µg/L (96 h)	Kumar et al. (2011b)
8	Danio rerio	0.119 µg/L (96 h)	Ansari and Ahmad (2010)
9	Channa punctatus	7.92 μg/L (96 h)	Kumar et al. (2007)
10	Cirrhinus mrigala	3.0 µg/L (96 h)	Velmurugan et al. (2007)
11	Brachydanio rerio	1.93 (Г) and 1.94 ( $\lambda$ ) µg/L (96 h)	Wang et al. (2007)
12	Brachydanio rerio	0.27 μg/L (96 h)	Sewell and McKenzie (2006)

of the fish, as a compensatory response to the circulating lymphocytes by the lymphoid tissues, and tissue damage (Ullah et al. 2019).

#### **Biochemical toxicity**

Biochemical parameters are often employed as handy biomarkers to appraise the toxic effects of different exogenous compounds, toxicants, and chemicals including pesticides, heavy metals, and pharmaceutical drugs on fish. Different generations of synthetic pyrethroids have been tested against different fish species, and almost all of them resulted in varying levels of biochemical toxicity, for example, deltamethrin induced different biochemical toxicities in *Hypophthalmichthys molitrix* including a marked reduction in the total protein contents in the liver, gills, muscles, blood, and brain tissues, marked increase in blood glucose concentration, and significant alterations in the concentration of potassium, sodium, chloride, total bilirubin, albumin, urea, inorganic phosphate, and cholesterol in serum (Ullah et al. 2019). Similarly, the activities of metabolic enzymes including aspartate

 Table 4
 Toxic effects of pyrethroids on fish as revealed by different biomarkers

S. no.	Fish species	Changes observed	References
Oxida	tive stress—ROS, LPO		
1	Hypophthalmichthys molitrix	Deltamethrin led to increased reactive oxygen species production and lipid peroxidation in the brain, liver, gills, and muscles. The antioxidant enzymes (catalase, peroxidase, superoxide dismutase, and glutathione reductase) activities were also increased in these tissues	Ullah et al. (2019)
2	Danio rerio	Deltamethrin induced oxidative stress leading to inhibition of acetylcholinesterase activity	Parlak (2018)
3	Prochilodus lineatus	$\lambda$ -Cyhalothrin induced lipid peroxidation in the liver, gills, brain, kidney, and muscle tissues along with an alteration in the activities of glutathione-S-transferase, superoxide dismutase, catalase, glutathione peroxidase, and glutathione concentration	Vieira and dos Reis Martinez (2018)
4	Cyprinus carpio	$\lambda$ -Cyhalothrin induced oxidative stress in terms of lipid peroxi- dase in liver, gills, and muscles and protein oxidation in the liver and muscles	Clasen et al. (2018)
5	Oncorhynchus mykiss	Cypermethrin exposure resulted in oxidative stress induction	Kutluyer et al. (2016)
6	Labeo rohita	Cypermethrin exposure resulted in increased LPO ( <i>thiobarbituric</i> acid) in gills, liver, brain, and muscles	Ullah (2015)
7	Oreochromis niloticus	Deltamethrin exposure led to oxidative stress in terms of an increase in malondialdehyde (LPO) in kidneys, liver, and gills with a decrease in the activities of catalase, superoxide dismutase, glutathione peroxidase, and glutathione contents in the tissues	Abdel-Daim et al. (2015)
8	Tor putitora	Cypermethrin induced lipid peroxidation in the brain, liver, mus- cles, and gills at acute concentration	Ullah et al. (2014)
9	Sparus aurata	Deltamethrin exposure led to oxidative stress induction leading to disturbed metabolism and immune status	Guardiola et al. (2014)
10	Cyprinus carpio	Deltamethrin increased malondialdehyde level in the hepatopan- creas with an increase in the activities of glutathione-S-trans- ferase, catalase, and glutathione reductase	Ensibi et al. (2013)
Develo	opmental toxicity		
1	Danio rerio	Bifenthrin altered the ryanodine and mTOR receptor-dependent signaling at the developing stage and also led to delayed hyper- activity	Frank et al. (2018)
2	Danio rerio	Deltamethrin exposure induced oxidative stress, apoptosis, and different morphological alterations at an embryonic stage	Parlak (2018)
3	Danio rerio	Deltamethrin exposure resulted in delayed development and sig- nificantly increased mortality at 24 and 48 h post-fertilization	Liu et al. (2018)
4	Labeo rohita	Cypermethrin induced developmental deformities and altered enzymatic activities in the developmental stages	Dawar et al. (2016)
5	Danio rerio	Bifenthrin and $\lambda$ -cyhalothrin disrupted hypothalamus-pituitary- thyroid axis at an embryonic stage	Tu et al. (2016)
6	Oncorhynchus mykiss	Bifenthrin disturbed the dopaminergic signaling at the juvenile stage	Crago and Schlenk (2015)
7	Danio rerio	Joint exposure to cypermethrin and permethrin induced different larval stage toxicities	Yang et al. (2014)
8	Danio rerio	Cypermethrin induced different morphological changes at embryo-larval stages	Shi et al. (2011)
9	Pimephales promelas	Bifenthrin changed the gene transcription at the larval stage	Beggel et al. (2011)
10	Danio rerio	Permethrin and deltamethrin induced different developmental neurotoxic effects such as craniofacial abnormalities, spasms, and body curvatures	DeMicco et al. (2010)
11	Danio rerio	Bifenthrin induced different morphological changes at embryo- larval stages	Jin et al. (2009)

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#### Table 4 (continued)

S. no.	Fish species	Changes observed	References
Neuro	toxicity		
1	Hypophthalmichthys molitrix	Deltamethrin inhibited acetylcholinesterase activity in the brain, muscle, and liver tissues and induced different behavioral inconsistencies such as erratic swimming, jumping, hyper- or hypo-activeness, increased air gulfing, rapid gill movement, and equilibrium loss	Ullah et al. (2019)
2	Alburnus tarichi	Cypermethrin mediated different changes in behavior such as hyperexcitability, balance loss, vertical position adaptation, rapid mouth open and gill movement, and irregular swimming, with an increase in cortisol (stress hormone)	Özok et al. (2018)
3	Channa punctatus	Deltamethrin inhibited acetylcholinesterase activity in the brain, muscles, and gills and led to different behavioral inconsisten- cies including hyperactivity, balance loss, elevated surface activity, convulsions, and increased rate of opercular activity	Singh et al. (2018b)
4	Danio rerio	Cypermethrin exposure resulted in significant loss of motor coordination and led to erratic swimming, frequent freezing behavior, and loss of the voluntary control	Nema and Bhargava (2018)
5	Jenynsia multidentata	Cypermethrin changed the acetylcholinesterase activity and altered the swimming behavior	Bonansea et al. (2016)
6	Oreochromis niloticus	Cypermethrin exposure altered the swimming pattern	Haque and Mondal (2016)
7	Danio rerio	Deltamethrin disturbed acetylcholinesterase activity and induced different behavioral toxicities	Ren et al. (2016)
8	Labeo rohita	Cypermethrin mediated various behavioral alterations such as jumping, adapting vertical position before death, hyper- and hypo-activity, and altered swimming pattern	Ullah (2015)
9	Tor putitora	Cypermethrin induced different behavioral changes including erratic swimming, hyper- or hypo-activeness, equilibrium loss, and increased air gulfing and operculum beats	Ullah et al. (2014)
10	Rhamdia quelen	Cypermethrin exposure leads to different behavioral alterations such as dyspnea, spiral swimming, swimming alteration, loss of balance, and upright swimming	Montanha et al. (2014)
11	Cyprinus carpio	$\lambda$ -Cyhalothrin inhibited acetylcholinesterase activity in the brain, liver, and muscle tissues at both acute and subacute concentration	Bibi et al. (2014)
12	Danio rerio	Fenvalerate induced apoptosis in the brain	Gu et al. (2010)
Hema	tological toxicity		
1	Hypophthalmichthys molitrix	Deltamethrin exposure altered hematological parameters includ- ing red blood cells, white blood cells, hemoglobin, hematocrit, packed cell volume, mean corpuscular volume, mean corpuscu- lar hemoglobin, and mean corpuscular hemoglobin concentra- tion	Ullah et al. (2019)
2	Alburnus tarichi	Cypermethrin mediated changes in hematology such as decreased red blood cells, hemoglobin, and hematocrit	Özok et al. (2018)
3	Prochilodus lineatus	$\lambda$ -Cyhalothrin altered red blood cells, white blood cells, and hemoglobin	Vieira and dos Reis Martinez (2018)
4	Cyprinus carpio	Permethrin decreased red blood cells, hemoglobin, and packed cell volume at both lethal and sublethal concentrations, whereas it increased white blood cells, mean corpuscular volume, mean corpuscular hemoglobin, and mean corpuscular hemoglobin concentration at both the concentrations	Gopala Rao et al. (2017)
5	Salmo trutta fario	Deltamethrin decreased white blood cells, hemoglobin, hema- tocrit, mean corpuscular hemoglobin, and mean corpuscular volume, while it increased the number of red blood cells	Karatas (2016)
6	Tor putitora	Cypermethrin induced changes in white blood cells and red blood cells	Ullah et al. (2015)

 Table 4 (continued)

S. no.	Fish species	Changes observed	References
7	Rhamdia quelen	Cypermethrin exposure altered erythrocytes, total leukocytes, hemoglobin, hematocrit, thrombocytes, neutrophils, monocytes, lymphocytes, basophil, and eosinophil	Montanha et al. (2014)
8	Catla catla	Cypermethrin decreased the hematological indices including hemoglobin, total leukocyte counts, total erythrocyte count, and hematocrit	Vani et al. (2012)
9	Catla catla	Deltamethrin disrupted hematopoiesis and led to a decrease in hematological parameters	Vani et al. (2011)
10	Labeo rohita	Fenvalerate affected the hematological parameters including red blood cells, white blood cells, hemoglobin, and hematocrit, mean corpuscular volume, mean corpuscular hemoglobin, mean corpuscular hemoglobin concentration, and nitroblue tetrazolium test	Prusty et al. (2011)
11	Cyprinus carpio	Bifenthrin induced different alterations in hematology	Velisek et al. (2009a)
Bioch	emical toxicity		
1	Hypophthalmichthys molitrix	Deltamethrin induced biochemical toxic effects by altering the protein contents in blood, liver, gills, muscles, and brain tissues; blood glucose, serum biochemistry such as sodium, potassium, chloride, total bilirubin, albumin, cholesterol, urea, and inorganic phosphate	Ullah et al. (2019)
2	Prochilodus lineatus	$\lambda$ -Cyhalothrin altered the concentration of glucose, sodium, potassium, chloride, calcium, and magnesium	Vieira and dos Reis Martinez (2018)
3	Brycon amazonicus	Cypermethrin increased branchial Na <sup>+</sup> /K <sup>+</sup> ATPase activity due to MRC proliferation because of increased cortisol	de Moraes et al. (2018)
4	Cyprinus carpio	$\lambda$ -Cyhalothrin exposure resulted in redox imbalanced and conse- quently led to different biochemical alterations such as inhibited acetylcholinesterase in the brain, and increased catalase and glutathione-s-transferase activities	Clasen et al. (2018)
5	Labeo rohita	Cypermethrin induced changes in total proteins and the activi- ties of different enzymes including superoxide dismutase, peroxidase, glutathione peroxidase, glutathione-S-transferase, catalase, and glutathione reductase	Ullah (2015)
6	Labeo rohita	Deltamethrin increased alkaline phosphatase and decreased acid phosphatase and acetylcholine esterase activities in the liver and kidney	Suvetha et al. (2015)
7	Tor putitora	Cypermethrin induced altered total protein contents, antioxidant enzymes activities, and blood glucose	Ullah et al. (2014)
8	Rhamdia quelen	Cypermethrin exposure mediated alterations in plasma protein concentration	Montanha et al. (2014)
9	Labeo rohita	Fenvalerate affected the activities of superoxide dismutase in liver and gills, and catalase in gills. Blood glucose level, serum creatinine, and triglycerides were increased, while serum total protein, globulin, and albumin were decreased	Prusty et al. (2011)
10	Oryzias latipes	Permethrin altered the concentration of vitellogenin protein in the liver	Nillos et al. (2010)
Physic	ological toxicity		
1	Hypophthalmichthys molitrix	Deltamethrin disturbed protein and glucose metabolism in brain, liver, gills, and muscles tissues	Ullah et al. (2019)
2	Alburnus tarichi	Cypermethrin significantly increased the activities of lactate dehydrogenase, alanine aminotransferase, and aspartate ami- notransferase	Özok et al. (2018)
3	Prochilodus lineatus	$\lambda$ -Cyhalothrin altered the activities of Ca <sup>+2</sup> ATPase, Mg <sup>+2</sup> ATPase, and Na <sup>+</sup> K <sup>+</sup> ATPase in the gills	Vieira and dos Reis Martinez (2018)
4	Carassius carassius	Deltamethrin disturbed the heart function in vitro by disturbing/ irregularities of the rhythm and rate of atrial beating, atrial contraction, electrical activity, and affecting sodium channels in the heart up to 48%	Haverinen and Vornanen (2016)

#### Table 4 (continued)

S. no.	Fish species	Changes observed	References
5	Labeo rohita	Cypermethrin exposure disturbed metabolic activities in brain, liver, gills, and muscles tissues	Ullah (2015)
6	Oreochromis niloticus	Deltamethrin disturbed the metabolic activities in the serum by altering the activities of alkaline phosphatase, aspartate ami- notransferase, and alanine aminotransferase	Abdel-Daim et al. (2015)
7	Rhamdia quelen	Cypermethrin exposure mediated alterations in the metabolic activities and altered the activities of alanine aminotransferase, aspartate aminotransferase, and alkaline phosphatase	Montanha et al. (2014)
8	Anabas testudineus	Permethrin reduced SDH level in the liver and muscle tis- sues, whereas deltamethrin decreased the level of aspartate aminotransferase in the liver and muscle tissues and alanine aminotransferase in the muscles	Sapana Devi and Gupta (2014)
9	Labeo rohita	Fenvalerate affected the activities of aspartate aminotransferase, alanine aminotransferase, alkaline phosphatase, glutamic pyru- vic transaminase, and glutamic oxaloacetic transaminase	Prusty et al. (2011)
10	Pimephales promelas	Bifenthrin decreased the transcription of the gene related to stress response, growth, and metabolism	Beggel et al. (2011)
Reproductive toxicity and endocrine disruptive toxicity			
1	Danio rerio	Bifenthrin induced alterations in the dopaminergic and estrogenic pathways at an embryo and larval stages	Bertotto et al. (2018)
2	Danio rerio	Permethrin, $\beta$ -cypermethrin, and their metabolites (PBCOH, PBCHO, and PBCOOH) altered the gene expression involved in thyroid and innate immune systems	Xu et al. (2018)
3	Oncorhynchus mykiss	Cypermethrin exposure resulted in the changed quality of sper- matozoa	Kutluyer et al. (2016)
4	Mugil cephalus	The mixture (mainly cyfluthrin and bifenthrin) altered the vitel- logenin genes expression in the liver of both female and male	Zheng et al. (2016)
5	Heteropneustes fossilis	Cypermethrin exposure changed the histoarchitecture of the ovary, e.g., cytoplasmic liquefaction, atretic follicle, granulosa layer invasion, ovarian wall degeneration, and inter-follicular spaces	Monir et al. (2016)
6	Danio rerio	Bifenthrin and $\lambda$ -cyhalothrin exposure resulted in thyroid endo- crine disruption	Tu et al. (2016)
7	Menidia beryllina	Bifenthrin disrupted estrogen signaling as well as decreased reproductive output in terms of fertilized egg/female	Brander et al. (2016)
8	Oncorhynchus mykiss	$\lambda$ -Cyhalothrin altered the quality and antioxidant responses of spermatozoa	Kutluyer et al. (2015)
9	Labeo rohita	Deltamethrin exposure resulted in an increased level of prolactin and plasma cortisol at both acute and subacute concentration	Suvetha et al. (2015)
10	Oncorhynchus mykiss	Bifenthrin increased ovarian follicle diameter and plasma estra- diol (E2) and reduced gonadosomatic index	Forsgren et al. (2013)
11	Menidia beryllina	Bifenthrin exposure resulted in an increased level of choriogenin (estrogen-responsive protein) at the juvenile stage	Brander et al. (2012)
12	Heteropneustes fossilis	Cypermethrin disturbed follicular wall and spermatogenic cells and induced changes in the plasma estradiol- $17\beta$ , sperm motil- ity, and 11-ketotestosterone	Singh and Singh (2008)
13	Salmo trutta	Cypermethrin exposure disturbed the reproductive behavior, such as less courting events, lower strippable milt volume, spending less time near females, and significantly lowered 11-ketotestos- terone	Jaensson et al. (2007)

#### Table 4 (continued) S. no. Fish species Changes observed References Histopathological-morphological toxicity 1 Hypophthalmichthys molitrix Deltamethrin exposure resulted in different histomorphological Ullah et al. (2019) alterations in the brain, e.g., hemorrhage, infiltration, neuronal degeneration, and spongiosis; liver, e.g., congestion, hemosiderosis, increased sinusoidal spaces, fibrosis, and pycnosis; gills, e.g., disruption of arch, epithelium, and lamellae, atrophy of lamellae, and desquamation of goblet cells; and intestine, e.g., shredding of mucosal cells, disruption of mucosal cells and intestinal mucosa, and increased goblet cells 2 Danio rerio Deltamethrin induced different developmental malformations Parlak (2018) 3 Deltamethrin induced different histopathological damages in the Colossoma macropomum Cunha et al. (2018) gills 4 Danio rerio Deltamethrin led to different morphological alterations such as Liu et al. (2018) shorter body length, larger head-body angle, and smaller eyes 5 Cypermethrin exposure induced different changes in the histo-Monir et al. (2015) Pangasianodon hypophthalmus logical architecture of gills and liver 6 Tor putitora Cypermethrin induced different histopathological damage to the Ullah et al. (2015) brain, liver, and gills 7 Clarias batrachus λ-Cyhalothrin induced different histo-architectural damages in Singh et al. (2015) the testes, liver, and kidneys 8 Sparus aurata Deltamethrin resulted in deleterious morphological changes in Guardiola et al. (2014) the liver 9 Aphanius dispar Deltamethrin exposure led to different histopathological changes Al-Ghanbousi et al. (2012) in the gills such as secondary lamellae fusion, epithelium lifting (lamellar), vacuolization, and desquamation 10 Oreochromis niloticus Deltamethrin mediated different histopathological lesions in the Kan et al. (2012) liver, e.g., necrosis, pycnosis, and hypertrophy, and gills, e.g., epithelial lifting, hypertrophy of mucous cells, and hyperplasia Bifenthrin induced different histopathological alterations 11 Oncorhynchus mykiss Velisek et al. (2009b) 12 Heteropneustes fossilis Cypermethrin induced different histological alterations in the Singh and Singh (2008) gonads, liver, and gonadotrophic cells 13 Oreochromis niloticus Deltamethrin exposure resulted in different histopathological Yildirim et al. (2006) lesions in the gills, e.g., fusion of secondary lamellae, gills hyperemia, and telangiectasis, and liver, e.g., hydropic degenerations Molecular toxicity: DNA damage/genotoxicity Hypophthalmichthys molitrix Ullah et al. (2019) 1 Deltamethrin damaged DNA in peripheral blood erythrocytes 2 P. lineatus λ-Cyhalothrin induced DNA damage in erythrocytes Vieira and dos Reis Martinez (2018) 3 Danio rerio Deltamethrin exposure leads to miss expression of ntl, shh, and Liu et al. (2018) krox20 4 Danio rerio Bifenthrin and $\lambda$ -cyhalothrin up-regulated the following genes: Tu et al. (2016) CRH, TTR, Dio2, UGT1ab, TRα, Pax8, and TSHβ 5 Menidia beryllina Bifenthrin down-regulated several estrogen-associated transcripts Brander et al. (2016) 6 Labeo rohita Cypermethrin induced DNA damage in blood cells Ullah (2015) 7 Prochilodus lineatus Cypermethrin induced in vivo DNA damage in the gills Poletta et al. (2013) 8 Oreochromis niloticus Deltamethrin mediated micronuclei formation in the erythrocytes Kan et al. (2012) 9 Gambusia affinis λ-Cyhalothrin exposure led to micronuclei induction and nuclear Gökalp Muranli and Güner (2011) abnormalities 10 Channa punctata Cypermethrin induced genotoxicity due to induction of reactive Ansari et al. (2011) oxygen species-mediated oxidative damage 11 Oncorhynchus mykiss Permethrin altered VTG-mRNA expression in hepatocytes Nillos et al. (2010) 12 Danio rerio Fenvalerate altered the expression of dlx2 and ogg1 genes Gu et al. (2010)

aminotransferases, alanine aminotransferases, lactate dehydrogenases, and glutamate dehydrogenases, and concentration of whole-body cortisol were significantly increased.

A number of well-documented studies revealed that different esters of synthetic pyrethroids induced different toxic impacts on the biochemical indices of various fish species, for example, permethrin altered the vitellogenin protein's concentration in the liver of Oryzias latipes (Nillos et al. 2010), fenvalerate increased blood glucose level, serum creatinine, and triglyceride and reduced total protein, globulin, and albumin in the serum of Labeo rohita (Prusty et al. 2011), cypermethrin decreased total proteins in the muscles, gills, brain, and liver and increased blood glucose in Tor putitora (Ullah et al. 2014), and deltamethrin increased alkaline phosphatase and decreased acid phosphatase in the liver and kidney of Labeo rohita (Suvetha et al. 2015). A number of such other changes have been reported for different fish species exposed to various synthetic pyrethroids such as cypermethrin-exposed Rhamdia quelen (Montanha et al. 2014) and Brycon amazonicus (de Moraes et al. 2018) and  $\lambda$ -cyhalothrin-exposed *Prochilodus lineatus* (Vieira and dos Reis Martinez 2018).

#### **Reproductive and endocrine disruptive toxicity**

The reproductive toxic effects and endocrine disrupting potential of synthetic pyrethroids are widely studied. They are known as endocrine disruptors, for example, they interfere with the receptors of steroid hormone, and exhibit anti-mineralocorticoid, anti-glucocorticoid, and anti-estrogenic effects (Zhang et al. 2016, 2018). Bifenthrin disrupted the development of testis, inhibited the sperm maturation, delayed spermatocyte development, and reduced testosterone and 17*β*-estradiol in Sebastiscus marmoratus (Li et al. 2017), decreased gonadosomatic index and increased ovarian follicle diameter and  $17\beta$ -estradiol in the plasma of Oncorhynchus mykiss (Forsgren et al. 2013), and significantly decreased the reproductive output of Menidia beryllina (Brander et al. 2016). Similarly, several studies documented different toxic effects of different pyrethroids on reproduction and endocrine disruption in different fish species, such as altering the dopaminergic and estrogenic pathways in Danio rerio (Bertotto et al. 2018), changing the spermatozoa quality in Oncorhynchus mykiss (Kutluyer et al. 2016), denaturing the structure of the ovaries in *Heteropneustes fossilis* (Monir et al. 2016), and up-regulating the vitellogenin gene expression in Oncorhynchus mykiss (Crago and Schlenk 2015), Pimephales promelas (Beggel et al. 2011), and Dario rerio (Jin et al. 2009).

#### Histomorphological and anatomical toxicity

Histopathological assessment in response to exogenous toxicants, environmental stressors, and abrupt deleterious environmental change is a powerful, useful, and key biomarker in ecotoxicological studies. It emerged as a key parameter in chemical risk assessment and safety studies using fish as a model organism because it is rapid and can be applied to a number of fish tissues such as kidneys, intestines, brain, gills, and liver. Moreover, it is a more sensitive biomarker than a single biochemical response because the histological changes reveal a transition of bioorganization from individual-level biochemical effect at a lower level to population-level effect at a higher level (Ullah et al. 2018a). For histopathological investigation, different important tissues of the fish are employed based on their significance and objective of the study. Gills are studied because of their involvement in different major functions including excretion, respiration, osmoregulation, acid-base balance, being primary contact organ to ambient water having the toxicants, and continuously exposed to the exogenous chemicals. Liver histopathology is often studied in aquatic toxicology because of being the detoxification center. The histopathological alterations in the intestine reveal typical stress induction in fish. Similarly, the histomorphological changes in the brain of fish can display a different level of severity, more specifically in response to synthetic pyrethroids because of their lipophilicity and efficient accumulative and absorptive capability of the fish brain.

A number of well-documented research studies demonstrated synthetic pyrethroids induced histomorphological alterations in different tissues of the exposed species of fish, for example, deltamethrin mediated different histopathological changes in the liver such as congestion, sinusoidal dilation, vacuolation, inflammatory cell accumulation, hemosiderosis, and cellular shrinkage, in the gills such as secondary lamellae folding, epithelium disruption, epithelium fusion, calcium accumulation, secondary lamellae detachment, secondary lamellae degeneration, and secondary lamellae fusion, in the brain such as spongiosis, neuronal degeneration, discoloration, and infiltration, and in the intestine such as disruption of mucosal cells, goblet cells increase, necrosis, and mucosal cells shredding of Hypophthalmichthys molitrix (Ullah et al. 2019). Similarly, a number of other histomorphological changes are observed in different tissues of different fish species in response to different synthetic pyrethroids, such as cypermethrin-exposed Tor putitora (Ullah et al. 2015) and Pangasianodon hypophthalmus (Monir et al. 2015), deltamethrin-exposed Oreochromis niloticus (Kan et al. 2012), Aphanius dispar (Al-Ghanbousi et al. 2012), Cyprinus carpio (Stará et al. 2015), Danio rerio (Parlak 2018), and Colossoma macropomum (Cunha et al.

2018), and bifenthrin-exposed *Oncorhynchus mykiss* (Velisek et al. 2009b).

#### **Molecular toxicity**

There is a growing body of emerging evidence, depicting the molecular toxicological impacts of synthetic pyrethroids on fish. Synthetic pyrethroids-induced DNA damage is well studied in different fish species, for example, cypermethrin induced genotoxicity in Channa punctatus (Ansari et al. 2011) and DNA damage in the gills of *Prochilodus lineatus* (Poletta et al. 2013) and in the erythrocyte of Labeo rohita (Ullah 2015) and  $\lambda$ -cyhalothrin induced DNA damage in the blood erythrocyte Prochilodus lineatus (Vieira and dos Reis Martinez 2018). Synthetic pyrethroids-mediated micronuclei induction is also well studied, for example,  $\lambda$ -cyhalothrin exposure led to nuclear abnormalities and induced micronuclei formation in Gambusia affinis (Gökalp Muranli and Güner 2011) and deltamethrin induced micronuclei formation in the erythrocyte of Oreochromis niloticus (Kan et al. 2012). There is also enough evidence regarding synthetic pyrethroids-mediated alterations in gene/mRNA expression, for example, permethrin altered VTG-mRNA expression in the hepatocytes of Oncorhynchus mykiss (Nillos et al. 2010). Moreover, research revealed that synthetic pyrethroids upregulate or down-regulate several transcripts or genes, for example, bifenthrin down-regulated several estrogen-associated transcripts in Menidia beryllina (Brander et al. 2016).

### Mechanism of action of synthetic pyrethroids

Synthetic pyrethroids adapt different mechanisms of toxicity; however, the primary mechanism is neurotoxicity or intoxicating the nervous system of the fish. Figure 2 shows a summary of neurotoxicity induction in response to synthetic pyrethroids and the leading subtle consequences. The schematic presentation of the mechanism of action of synthetic pyrethroids is provided in Fig. 3. Synthetic pyrethroids such as cypermethrin form cyanohydrin, which is decomposed into aldehydes and cyanides and subsequently results in the production of reactive oxygen species (Ullah et al. 2018b). Reactive oxygen species induce lipid peroxidation, increase oxidative stress leading to oxidative damage, and increase the concentration of calcium in the cytosol which in turn leads to cytotoxicity and genotoxicity in fish (Ullah 2015). Synthetic pyrethroids layer on the nerve cells and hinder the sodium channels during repolarization, which lead to an unconstrained depolarization and disturbed transmission of the driving forces. The adverse impacts of the synthetic pyrethroids are mainly attributed to their neurotoxic effects linked with the pathological retention of acetylcholine in the synaptic gaps and inhibition of acetylcholinesterase, giving rise to multiple nerve impulses and consequently leading to decreased cholinergic transmission. Moreover, exposure to synthetic pyrethroids results in trans-activated p53 leading to induction of MiR-200 and consequently resulting in apoptosis. Similarly, synthetic pyrethroids change the mitochondrial proteome, leading to mitochondrial dysfunction and subsequently leading to apoptosis, whereas the induced oxidative stress ultimately results in nigrostriatal dopaminergic neurodegeneration (Ullah et al. 2018b).

# Control and prevention of synthetic pyrethroids

In order to minimize the use of synthetic pyrethroids, an adaptation of proper strategies and practicing proper management should be ensured. Synthetic pesticides should be phased out gradually and continuously till completely phased out. In order to reduce potential risks, ecological farming should be adopted instead of industrial agriculture. Multi-level approaches should be adapted for crops protection, rather than exclusively depending on pesticides. This will elevate landscape heterogeneity, increase suitable habitats for pollinators, and control pests naturally or biologically. Vegetation should be actively managed, which will increase functional biodiversity. Crops should be types-wise and cultivar-wise rotated to increase soil fertility and make crops resistant to the pest. Natural agents should be used for bio-control such as the introduction of beneficial insects, viruses, bacteria, and nematodes. This will also improve crop protection (Douglas and Tooker 2015).

## Control and preventive measures against pesticides

Pesticides should be employed according to the regulation and should be used by following the stipulated regulations. Pesticides risk assessment and safety, biopesticides use in agriculture, and biotechnological advancement of agriculture should be exclusively included in future plans. The advanced form of constructed wetlands should be employed, which emerged as a more reliable management approach and treatment system for alleviating different nonpoint sources of pesticides including agricultural runoffs and draining. Through this advance system of wetlands, pesticides are evacuated via different processes such as biological processes including plant absorption or metabolism, physical processes including absorption, sedimentation, co-precipitation, and precipitation, chemical processes, e.g., hydrolysis, photolysis, cation exchange, oxidation, and reduction, and biochemical processes such as microbial deprivation (Vymazal and Březinová 2015).

Pesticides use should be avoided and restricted at local and home level via using lesser or no cosmetics. Biological pest management should be adapted. Pesticides should be locked and stored in childproof containers, cupboards, or cabinets. Pesticides with the least hazardous impacts and dangers should be used. The pesticides users, exposed masses, applicators, dealers, and farmers should be educated properly. They should be guided regarding manufacturers' suggestions, instructions, protection equipment, and avoiding exposure of pregnant women, infants, toddlers, and children to the pesticides. Similarly, at the community level, organic farming and integrated pest management should be adopted at public buildings such as schools, hospitals, and public parks and mass awareness programs including workshops, seminars, and symposia should be arranged. The government should instruct children, pesticides dealers, pesticides applicators, pesticides users, and general masses about the hostile effects of pesticides, at the national level. The environmental protection and public health organizations should monitor and regularly assess pesticides concentration in the local environmental media. They should restrict the use of illegal and banned pesticides. Regular pesticide-based poisoning surveillance and epidemiological studies should be a part of their plan. Permissible limits for the pesticides should be established. These organizations should try to restrict pesticide use within their defined limits and should establish pesticide poisoning control and emergency center.

### Conclusion

Pyrethroids have been reported from the soil, cash crops, land or terrestrial organisms, water, sediments, and aquatic organisms including fish. Therefore, it is threatening at a fish biodiversity standpoint, as pyrethroids implicated population decline of fish has been confirmed by various studies in the past. Moreover, different aquatic- and ecotoxicological studies revealed the severe toxic effects of synthetic pyrethroids on fish at various biological levels such as at molecular, cellular, histological, organismal, and population level. These studies provide a future window for further studies. To comprehensively appraise the hostile impacts of synthetic pyrethroids and explain the underlying mechanism more deeply, studies that can possibly link these different levels of biological impacts are highly recommended. Furthermore, toxicological studies regarding individual enantiomers of the pyrethroids should be undertaken. The knowledge from such experiments that are based on the enantioselective toxicity and chirality of the pyrethroids will help in developing environment-friendly pyrethroids. This will also enrich activity

of pyrethroids against target insects without posing severe hostilities on nontarget organisms including fish.

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